**Pinus sylvestris** L. subsp. *nevadensis* (Christ) Heywood in southern Spain: An endangered endemic Mediterranean forest

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**Abstract**

*Pinus sylvestris* subsp. *nevadensis* is the most endangered pine in Spain. This taxon takes refuge in only two massifs of the Betic Cordillera –Sierra Nevada and Sierra de Baza–, where its forests represent the southernmost limit of the species global distribution, surviving under conditions of geographic, demographic and ecological marginality in the upper treeline boundary that makes them very vulnerable to any environmental change or external aggression. This research establishes for the first time, and for the entire Betic Cordillera, the locations and ecological patterns of these pine forests, their plant dynamic and floristic composition, and provides an updated map of the current and potential distribution area of this subspecies. The methodological process for this research has consisted of an integrated phytosociological and biogeographical analysis of vegetation and the resulting landscape, through fieldwork covering in as much detail as possible the distribution area of *P. sylvestris* in the Betic Cordillera, and a review of the bibliographic background. Taking into account the results, *P. sylvestris* subsp. *nevadensis* forests survive at present under hostile Mediterranean conditions due to the special physical characteristics of the microenvironments in which they have taken refuge, mainly the cool, relatively moist climate of their ecological niches and the relatively impermeable soils that forests occupy. However, there are significant ecological obstacles for the future preservation of this pine in southern Spain, and therefore it is necessary for the creation of programmes to protect these threatened endemic and post-glacial relict forests through continued monitoring of their evolution and further research studying the processes that make this ecosystem as a whole so unique and valuable.

**Keywords:** Scot pine, Sierra Nevada, Sierra de Baza, ecological patterns, plant dynamic, floristic composition, current distribution map.

**Introduction**

Scot pine – *Pinus sylvestris* L.– is the most widely distributed taxon of the *Pinus* genus around the world, occupying an outstanding place in the European vegetation throughout the Quaternary period (the last 1.6 Ma) (Kelly, Connolly, 2000; MacAllister, 2016). The species has an extensive geographical distribution in much of Europe and Asia at present – it is the most common tree in these territories– having been cultivated for a long time (Koski, 2000), which colonizes numerous mountain habitats covering a broad latitudinal gradient of ecological conditions, therefore having an important morphological diversity (Costa, et al., 2005; MacAllister, 2016). The native forests grow in wet-cold, very cold, and extremely cold continental bioclimates, preferably in oroboreal phytoclimates (Rosúa, et al., 2001; Costa, et al., 2005) of the Eastern-Siberian, Euro-Siberian, and Mediterranean bioge-
graphic regions of the Holarctic Kingdom. This is the widest natural chorological area of any pine species on the Earth (Kelly, Connolly, 2000; Mason, Alía, 2000).

South of its ecological area, *P. sylvestris* is an increasingly rare taxon, particularly in the Mediterranean Basin, where the species has a terminal distribution conditioned by the dry boundary of the sub-Mediterranean climate, remaining confined within this climate or in other xericer temperate or sub-xericer cold domains (Ruiz de la Torre, 2006). The availability of edaphic and environment humidity during the summer under Mediterranean conditions provides its main limiting factor (Xenakis, et al., 2012). For example, it has been found that the specific leaf area is lowest for the Spanish provenances of *P. sylvestris* in comparison with other Eurosiberian provenances, which is related to the droughts and dry periods (MacAllister, 2016). In particular, in the Iberian Peninsula these forests took refuge in certain high mountain regions after their regression since the end of the Last Glacial Maximum (Ruiz-Labourdette, et al., 2012), where they currently form the western and southernmost limits of the natural distribution range of *P. sylvestris* (Mason, Alía, 2000; Andreu, et al., 2007; Martínez-Vilalta, et al., 2008). These circumstances have favoured its disjunct distribution in several Spanish mountain ranges and the development of five subspecies derived from two different Iberian haplotypes that existed during the Würm period (Cheddadi, et al., 2006): *catalaunica* Gaussen and olivicola Vayr (NE Iberian mountains), *iberica* Svoboda (Central and Iberian Cordilleras), *pyrenaica* Svoboda (central and western Pyrenees), and nevadensis (Christ) Heywood (Betic Cordillera, southern Iberian Peninsula).

Of all these Iberian subspecies, *P. sylvestris* subsp. nevadensis is the most endangered occurring only in two geographically isolated massifs of Betic Cordillera (southern Iberian Peninsula): Sierra Nevada (National Park and Biosphere Reserve) and Sierra de Baza (National Park). In these massifs, the subspecies is found between 1,600 and 2,200 meters above sea level, occupying an ecological boundary of an upper tree-line that makes it very vulnerable to any environmental change or external aggression. The Mediterranean climate (especially the resultant stress due to summer dryness), the small size of forests, and the lack of regeneration due to overgrazing are significant ecological obstacles to its preservation (Mason, Alía, 2000). Therefore, the subsp. nevadensis forests are considered endangered relict formations of great interest (Ruiz de la Torre, 2006) due to their paleoenedemar character (based on palinological studies), high ecological value, and extreme fragility (Olmedo-Cobo, 2011).

This research presents for the first time a geoeological analysis of *P. sylvestris* subsp. nevadensis for the entire Betic Cordillera. Together with a review of earlier primary studies of this taxon (Valle, Gómez, 1988; Pérez-Raya, et al., 1990; Blanca, Morales, 1991; Molero, et al., 1992; Castro, 1999; Zamora, et al., 2001; Castro, et al., 2004, 2005; Costa, et al., 2005; Ruiz de la Torre, 2006; Olmedo-Cobo, 2011; Olmedo-Cobo, et al., 2014), fieldwork has been the main tool of analysis. The results obtained determine the current circumstances in which the species develops in such a hostile Mediterranean environment as the Betic Cordillera, and provide an updated map of the area occupied by these forests at present, corroborating and, where appropriate, reviewing the existing information. The data collected provides a valuable and integral tool that can be used for forestry management and the future preservation of this species, particularly in the face of the threats posed by climate change.

### Methods and data

#### Biogeography and bioclimatic context of study area

The massifs of Sierra Nevada (37°03′12″N, 3°18′41″W, Mulhacén peak-3,481 meters above sea level) and Sierra de Baza (37°22′50″N, 2°50′31″W, Santa Bárbara peak-2,269 m asl) are located in the central Betic Cordillera (southern Iberian Peninsula, Spain) (Figure 1), with a respectively territorial extension of 2,000 km² and 600 km².

These Holarctic territories belong to the Betic biogeographic province of the Mediterranean region (western Mediterranean subregion), and according to the latest biogeographical zoning of the Betic province, carried out by Rivas-Martínez (2011), the *P. sylvestris* subsp. nevadensis forests are located in the Trevenquino district of the Granadino-Almi-jarense sector in Sierra Nevada, and in the Serrano-Bastetano district of the Accitano-Bastetano sector in Sierra de Baza. Based on the premises established by Allué (1990), Rivas-Martínez and Loidi (1999), Rivas-Martínez et al. (2004) and Rivas-Martínez (2008) for the Mediterranean region, the *P. sylvestris* forests occur inside the subhumid-humid Supra-Oromediterranean bioclimatic in Sierra Nevada, where the annual average rainfall is estimated to be 800-1,000 mm and the annual mean temperature ranges between 6°C at 2,000 m asl and 8°C at 1,600 m asl. Comparatively, the bioclimatic optimum of the nevadensis subspecies in Sierra de Baza corresponds to the subhumid Oromediterranean zone, with an annual average rainfall of approximately 700-800 mm and an annual mean temperature of 5-7°C above 1,800 m asl according to Olmedo-Cobo (2011).
Methodology

The methodological process has been based on an integrated phytosociological and biogeographical analysis of vegetation and of the resulting landscape, which involves the study of the habitats, the vegetation series and geo-series, and the communities and floristic associations that comprise the dynamic and the plant succession of subsp. *nevadensis* forests. Through an approach based on extensive fieldwork the study area has been covered in as much detail as possible. This systematic review has established: (1) the ecological patterns of *P. sylvestris* subsp. *nevadensis* distribution in Sierra Nevada and Sierra de Baza massifs, (2) the plant dynamics and floristic composition of the different phytocoenosis recognized, and (3) the potential and current area –mapping– of *P. sylvestris* subsp. *nevadensis* forests.

The main references used for the designation, authorship, and verification of vegetation series and flora have been the works of Valle and Gómez (1988), Blanca and Morales (1991), Molero et al. (1992), Valle (2003), Castro et al. (2004; 2005), Castroviejo (2010) and Blanca et al. (2011). The study of the floristic composition of forests has been based on the Sigmatist Methodology of the Zurich-Montpellier school and the Braun-Blanquet abundance-cover index1 (see Braun-Blanquet, 1979). GIS (ArcGis 10.1) has been used for mapping, on the basis of data collected through fieldwork and the examination of aerial photographs from 2011, but also from 1977 and 1956 for the differentiation between native forests and reforestation masses. The lithological and edaphic references utilised to check the collected data have been obtained from Vera (2004), and the soils maps of the Lucdeme Project (scale 1:100,000) (Spanish Ministry of Agriculture, Fisheries and Food, 1988-1990), respectively.

Results

Locations and distribution patterns

*P. sylvestris* subsp. *nevadensis* in Sierra Nevada is located in the Monachil River basin near the Trevenque and Tesoro peaks between 1,600 and 2,000 m asl (west of the massif, only 20 km of Granada city). The main forest cores occur in the N-NE slopes of Pico del Tesoro, in the confluence between the headwaters of the Benalcázar river and the Matas Verdes Hill (figure 2 left), and in the Huenes river near to Trevenque peak (E-NE slope of Trevenque peak). In Sierra de Baza (figure 2 right), the forests grow on several of the central and western slopes, between 1,800 and 2,250 m asl, mainly around the famous enclave of the natural park, Prados del Rey. Specifically, the most developed pinewoods appear on the W, NW, N, and NE slopes of Santa Barbara peak, on the N and NE slopes of the Pozo de la Nieve peak, on the N and NW slopes of the

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1 Abundance-cover index: 5, any number of individuals, coverage > 75% of the study area; 4, any number of individuals, coverage 50-75%; 3, any number of individuals, coverage 25-50%; 2, any number of individuals, coverage 5-25%; 1, numerous individuals, coverage < 5%; 0, scattered individuals, coverage > 5%; +, few individuals, little coverage; R, isolated individuals, little coverage.
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Cerro del Calar and Cerro del Sabinar peaks, and locally on the NW, N, and NE slopes of the San Sebastian and Rapa peaks; also, there are a few isolated individuals on other south orientations, as on the Peñón de la Cogollada peak, on the Relumbre Ravine, and on the La Boleta and Los Tejoletos hills.

At these locations, *P. sylvestris* occupies not overly steep slopes, oriented to north, northwest and northeast, emplacements that have an optimal exposure to the moist westerly air masses –associated with the polar jet stream– and a temperate (not warm) thermal regime in summer due to the elevation. The forests are also relatively protected from the scouring action of winds, which become an absolute limiting factor for woodlands on the summits of both massifs. Locally, the pines grow on steep periglacial screes, where water reaches the more superficial rocky strata, being usable as reservoir by the trees during the summer or during drought periods. These are conditions that favour a moist, shady microclimatic environment with lower evaporation and evapotranspiration values and less sunshine during the hot and dry season with regard to the lower areas of both massifs.

Following the topographic and microclimatic conditions, the edaphic factor is the other essential issue that explains the current persistence of *P. sylvestris* in the Betic mountains. Although the biogeographical districts where the nevadensis subspecies appears have a predominance of very rocky and not overly developed basic lithosols over calcareous substrates –mainly Triassic limestones and dolomites–, its forests mostly occupy more evolved acidic soils –of eutric regosol type– that appear locally on siliceous substrates such as calc-schists, quartizes, and phyllites. In addition to the clear contrast in the degree of evolution between these basic and acidic edaphic domains, there is a remarkable difference in the ability to retain moisture between them. Compared to the high capacity of the basic soils for water percolation, which causes significant edaphic xericity, the acidic soils are relatively impermeable and even experience frequent flooding phenomena during the rainy season (autumn and winter) and spring thaw. Thus, in the most xeric enclaves, that is, in the rockiest soils of a carbonate nature and also in locations with less favourable microclimates (such as on the steepest slopes and on some of the highest, most storm-beaten peaks), *P. sylvestris* disappears quickly, especially in southern orientations, being gradually replaced by *P. nigra* Arnold subsp. *latisquama* (Willk.) Heywood in Oromediterranean level, and by *P. pinaster* Aiton, *Quercus rotundifolia* L., and *Acer gramineum* Boiss. in Supramediterranean area. *P. sylvestris* also disappears in enclaves where the almost permanent seasonal soil waterlogging allows only the development of hygrophilous grasslands of *Festuca iberica* (Hack.) K. Richt., which are considered out of the plant succession of the forest.

**Plant dynamic and floristic composition of *P. sylvestris* subsp. nevadensis phytocoenosis**

The forests of *P. sylvestris* subsp. *nevadensis* integrate the *Daphne hispanicae-Pinetum nevadensis* association, which constitute the climax stage of the vegetation series *Daphno hispanicae-Pinetonevadensis Sigmetum*. This one is defined by Rivas-Martinez (2011) as the series of the forests of *P. sylvestris* subsp. *nevadensis* and *Daphne oleoides* Schreb. with *Juniperus sabina* L. and *P. nigra* subsp. *latisquama*, of limestone and dolomite nature, climatophilous and edaphoxerophilous character, and endemic to Sierra Nevada and Sierra de Baza. The climax community occurs on limestone-dolomite rocks and, although its bioclimatic optimum is the Oromediterranean zone, the series also appears in Supramediterranean enclaves due to topographical reasons that prevent the develop-
ment of vegetation typical of this level, as Molero et al. (1992) have identified.

The climax stage is characterized by forests of heliophilous nature and simple stratification, denser on more humid, cooler slopes over siliceous substrates. In rock enclaves, the forest has a progressively more open configuration as a result of greater soil xericity, which causes the replacement of *P. sylvestris* by *P. nigra* subsp. *latisquama*, with a very limited coexistence of both species, which only occurs locally in Sierra de Baza. There is a predominance of mature trees, some very long-lived, reaching heights of more than 12 m, and relatively few young individuals and seedlings. In addition to the forests, the climax is completed with creeping thickets of junipers of *J. sabina* (figure 3 left above) and creeping junipers of *J. communis* L. subsp. *hemisphaerica* (K. Presl) Nyman, which are scattered under the trees and on sparsely wooded or un-wooded slopes indistinctly, where they colonize even the rockiest soils with variable coverage. Phytosociological inventories show in detail the structure and floristic composition of the *Daphno hispanicae-Pinetum nevadensis* community (Table 1).

However, the marginal ecological boundary where the *P. sylvestris* subsp. *nevadensis* forests grow induce a plant ecosystem with a seriously degraded structure, characterized by serial vegetation stages, with a characteristic flora cohort (Table 2), that frequently replaces the climax communities:

Table 1. Structure and floristic composition of *Daphno hispanicae-Pinetum nevadensis* association (phytosociological inventories)

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<td>191</td>
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<td>Degree of slope (%)</td>
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Association characteristic taxa

| *Pinus sylvestris nevadensis* | 4 | 3 | 2 | 3 | 1 | 4 | 3 | 3 | 3 | 1 |
| *Pinus nigra latisquama* | . | . | 3 | 1 | + | . | . | . | . | . | . | . |
| *Juniperus sabina* | 2 | 2 | 2 | 1 | 4 | 3 | 2 | 2 | 4 | 2 | 1 | 3 |
| *Juniperus hemisphaerica* | 1 | r | 1 | . | + | 2 | 1 | 1 | + | + | 1 | 3 |
| *Daphne hispanica* | . | r | . | r | r | r | 1 | + | . | . | + | + |
| *Poa ligulata* | . | + | . | . | + | . | . | . | . | . | . | . |

Companion taxa

| *Acer opalus granatensis* | + | . | + | . | . | . | + | . | . | . | + | + |
| *Berberis hispanica* | 2 | 1 | 2 | 1 | 1 | 2 | 1 | 2 | 1 | 1 | 2 | 1 |
| *Vella spinosa* | 1 | + | 1 | 1 | + | . | + | . | 1 | 1 | r | + |
| *Hormathophylla spinosa* | 1 | + | 1 | r | + | + | + | . | + | . | + | + |
| *Rosa sícula* | + | 1 | 1 | . | + | 1 | + | r | . | + | . | + |
| *Erinacea anthyllis* | . | + | . | + | . | + | . | + | 1 | . | . | + |
| *Astragalus granatensis* | + | + | 1 | . | + | 1 | r | 1 | 1 | . | . | + |
| *Rosa pouzinii* | r | . | . | . | . | . | 1 | + | . | . | . | . |
| *Prunus ramburii* | + | . | . | + | . | + | . | r | . | + | . | r |
| *Prunus prostrata* | + | . | . | + | . | + | . | r | . | + | . | + |
| *Ononis aragonensis* | + | . | r | . | r | 1 | 1 | 1 | + | . | + | + |
| *Amelanchier ovalis* | . | r | + | . | . | . | + | + | r | . | . | + |
| *Crataegus monogyna* | . | . | + | r | . | r | 1 | 2 | . | 2 | 1 | . |
| *Lonicera arborea* | . | r | . | . | r | . | . | . | + | . | . | . |
| *Lonicera splendida* | . | . | . | . | . | . | + | . | . | . | . | . |
| *Paonia officinalis* | . | . | . | . | . | . | . | r | . | . | . | . |
| *Hormathophylla longicaulis* | . | . | . | . | . | . | + | . | . | . | . | . |
| *Arenaria armerina* | . | . | . | . | . | . | 1 | . | . | + | + | . |
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*Source: Authors (own elaboration)*

**Locations:** Sierra de Baza: 1. Prados del Rey (37°22’41"N, 2°51’18"W); 2. Cerro del Calar (37°22’21"N, 2º51’13"W); 3. Pinar de la Fonfría (37°22’15"N, 2º52’30"W); 4. Collado de la Boleta (37°22’53"N, 2º51’43"W); 5. Calar de Casa Heredia (37°21’59"N, 2°49’39"W); 6. Collado del Sabinar (37º22’49”N, 2º51’46”W).

Sierra Nevada: 7. Arroyo de Huenes (37º04’55”N, 3°28’08”W); 8. Pico del Tesoro (37º5’35”N, 3°27’23”W); 9. Monte Ahi de Cara (37º7’33”N, 3°25’54”W); 10. Arroyo de Benalcázar (37º05’17”N, 3°27’07”W); 11. Arroyo de Huenes (37º04’45”N, 3°28’17”W); 12. Pico del Tesoro (37º5’28”N, 3°27’30”W).

**Table 2.** Main flora of the climax community and of the serial stages that integrate the plant succession of *P. sylvestris* subsp. *nevadensis* forests

<table>
<thead>
<tr>
<th>Vegetal stage</th>
<th>Associations and main taxa</th>
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<tr>
<td><strong>Preforestal thorny belt</strong></td>
<td><strong>Lonicero-Berberidetum hispanicae</strong> sub-ass. <strong>juniperetosum sabinae.</strong> Characteristic species: <em>Berberis hispanica</em>, <em>Amelanchier ovalis</em>, <em>Crataegus monogyna</em>, <em>C. granatensis</em>, <em>Cotoneaster granatensis</em>, <em>Lonicera arborea</em>, <em>L. splendida</em>, <em>Prunus prostrata</em>, <em>P. ramburii</em>, <em>Rhamnus saxatilis</em>, <em>Rosa stylosa</em>, <em>R. pouzinii</em>, <em>R. sicula</em>, <em>R. micrantha</em> and <em>Ononis aragonensis</em>.</td>
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<td><em>Source: Based on the fieldwork and previous studies of Valle and Gómez (1988), Blanca and Morales (1991), Valle (2003) and Olmedo-Cobo (2011)</em></td>
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**Serial grasslands**

| **Seselio-Festucetum hystricis.** Characteristic species: *Seseli montanum* subsp. *granatense*, *Festuca hystricis*, *Poa ligulata*, *Flaccidula*, *A. armerina*, *Koeleria vallesiana*, *Jurinea humilis*, *Odontites viscosus* subsp. *granatensis* and *Alyssum serpyllifolium*. |

**Source:** Based on the fieldwork and previous studies of Valle and Gómez (1988), Blanca and Morales (1991), Valle (2003) and Olmedo-Cobo (2011)
Preforestal belt. The first serial stage corresponds to thorny thickets of usually modest height (0.5 to 1.5 m) and variable coverage, which occupy the peripheral areas around forests and clearings inside them (figure 3 right above). This stratum is progressively less dense on the rockier, steeper slopes due to greater soil xericity. This community belongs to the *Lonicero splendidae-Berberidetum hispanicae* association which, although it has a Supramediterranean optimum, also grows in the lower horizon of the Oromediterranean level, where it is designated as *juniperetosum sabinae* sub-association (junipers and creeping junipers differentiate it from the Supramediterranean typical stage).

Xeroacanthic thickets. Spiny formations well adapted to high-mountain mesological conditions. These communities grow mostly in the driest, rockiest, and most eroded soils and in places very exposed to wind and solar radiation, where they replace the forests, creeping thickets, and preforestal thorny belts (figure 3 left bottom). This stage has a different taxonomic classification in the Granadino-Almijarense and Accitano-Bastetano sectors: the community is known as *Festuco hystricis-Astragaletum granatensis* in Sierra Nevada, while in Sierra de Baza it is classified as *Saturejo intricatae-Velletum spinosae* association. Finally, endemic thyme formations of small size and open cover appear locally located over sandy soils (with superficial fragmentation of dolomitic rocks), belonging to the *Convolvulo nitidi-Andryaletum agardhi* association.

Serial grasslands. Two main stages of serial grasslands can be differentiated in the plant succession of *Daphno-Pinetum nevadensis* depending on the development and moisture-holding capacity of the soils. First, Oromediterranean grasslands of a perennial character (*Coronillo minimae-Astragaletum nummuraloides*) develop in the most evolved basic soils, which hold moisture in spring and until the beginning of the dry season (parching during the summer), especially in topographically and geomorphologically favourable enclaves such as karst depressions (figure 3 right bottom). The second formation corresponds to

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**Figure 3.** Creeping thickets of *J. sabina* colonizing the most rocky and xerophytic slopes (Santa Bárbara peak, Sierra de Baza) (left above); Preforestal thorny belt of the *P. sylvestris* subsp. nevadensis forests in Collado de Matas Verdes (Sierra Nevada) (left right); Spiny scrub of *Hormathophylla spinosa* (L.) P. Kupfer, *Erinacea anthyllis* Link ssp. *anthyllis* and *Vella spinosa* Boiss. in Calar de Rapa (Sierra de Baza) (left bottom); Serial grassland of *Coronillo minimae-Astragaletum nummuraloides* near the Tesoro peak (Sierra Nevada) (right bottom)

Source: Olmedo-Cobo
Pinus sylvestris L. subsp. nevadensis (Christ) Heywood in southern Spain: An endangered endemic Mediterranean forest

grasslands with small grasses and low soil coverage, which grow in rocky enclaves of high xericity (Seselio granatensis-Festucetum hystricis).

*P. sylvestris* subsp. *nevadensis* forests comprise the upper wooded strip in the vegetation geoseries of the northern slopes of Sierra Nevada and Sierra de Baza (Figure 4 AB), and constitute the highest treeline boundary in the Betic Cordillera. These pinewoods can be considered a transitional woodland domain between the temperate forests occupying the middle Betic mountain [Mesomediterranean and Supramediterranean sclerophyllous and coniferous forests of *Q. rotundifolia*, *Q. Faginea* Lam., *Q. Pyrenaica* Willd., *P. pinaster*, *P. halepensis* Mill. and *P. nigra*, and locally Supramediterranean deciduous forests of *A. monspessulanan L.*, *A. opalus* subsp. *granatensis*, *Sorbus aria*

![Figure 4](image-url)

**Figure 4.** Position of *P. sylvestris* subsp. *nevadensis* forests in the vegetation geoseries of the northern slope of (A) Sierra Nevada and (B) Sierra de Baza.

Source: Authors

(A): 0, Anthropogenic areas without vegetation. 1, Climatophilous forests of *Quercus rotundifolia*. 2, Edaphohygrophilous riparian formations of *Salix* and *Populus*. 3, Climatophilous forests of *Q. rotundifolia*. 4, Edapheroxerophilous forests of *P. pinaster*. 5, Climatophilous forests of *Acer granatensis*. 6, Climatophilous forests of *P. sylvestris* subsp. *nevadensis*. 7, Climatophilous forests of *Q. rotundifolia*. 8, Edaphohygrophilous riparian formations of *Carex* and *Salix*. 9, Climatophilous forests of *Q. pyrenaica*. 10, Creeping and xeroacanthic scrubs. 11, Psicroxerophilous grasslands. 12, Edaphohygrophilous geopermaseries.

(B): 0, Anthropogenic areas without vegetation. 1, Edapheroxerophilous forests of *P. halepensis*. 2, Climatophilous forests of *Q. rotundifolia*. 3, Edaphohygrophilous riparian formations of *Salix* and *Populus*. 4, Sub-spontaneous pinewoods of *P. halepensis*. 5, Climatophilous forests of *Q. rotundifolia*. 6, Edaphoxerophilous forests of *P. halepensis*. 7, Climatophilous forests of *A. granatensis*. 8, Edaphoxerophilous forests of *P. latisquama*. 9, Climatophilous forests of *P. sylvestris* subsp. *nevadensis*. 10, Edaphohygrophilous geopermaseries.

**Cartography of current and potential area of forests**

The potential distribution area of *P. sylvestris* subsp. *nevadensis* reaches 13.32 km² in Sierra Nevada and 29.71 km² in Sierra de Baza (Figure 5). In both cases, it is difficult to calculate accurately how much of the phytocoenosis domain is occupied at present by indigenous pinewoods due to the reforestation carried out in the 1950s and 1960s, which has greatly denaturalized the plant ecosystem. As a result, the coverage of the potential surface area by natural forests at present is minimum: it has been established that only 2.17 km² in Sierra Nevada (localized on the south-western core of the subspecies *nevadensis* potential area, Figure 6A) and 5.12 km² in Sierra de Baza (Figure 6B) are forests with an arboreal cover of adult specimens mostly.

![Figure 5. Potential distribution area of *P. sylvestris* subsp. *nevadensis* in Sierra Nevada and Sierra de Baza](source)

*Source: Based on fieldwork and modified from Valle and Gómez (1988), Valle (2003) and Olmedo-Cobo (2011)*

![Figure 6. Current surface occupied by indigenous forests of *P. sylvestris* subsp. *nevadensis* in Sierra Nevada (A) and Sierra de Baza (B)](source)

*Source: Based on fieldwork and modified from Valle and Gómez (1988), Valle (2003) and Olmedo-Cobo (2011)*
native (>60/80%). The rest of the potential area is occupied by afforestation masses of *Pinus (sylvestris, nigra, pinaster)* or by replacement stages of the climax, among which stands out the scrubs of *J. sabina* and *J. communis* due to its ecological role of providing nurse formations for pine seedlings.

**Discussion**

The natural regression related to climate evolution after the last glacial maximum and human-induced degradation of the *P. sylvestris* subsp. *nevadensis* forests by agriculture, overgrazing, mining, and forestry in the Betic Cordillera during the last centuries is reality. The presence of refugees of this species in two southern Iberian Peninsula massifs is undoubtedly related to the existence of benign microenvironments in these mountains, which may lead to the species being included in the group of European tree species whose local climatic shelter at their range limit is still associated with glacial refuges (Tzedakis et al., 2002; Svenning, Skov, 2007). As pointed out by Linares (2011) for the Mediterranean firs (*Abies* genus), the orographical complexity of the Mediterranean region has presumably also played a key role in the evolution and persistence of *P. sylvestris* forests in southern phyto-refuges, such as the Betic Cordillera, after the last glaciation.

From the data collected during this research, the main key to the persistence of *P. sylvestris* subsp. *nevadensis* seems to be the edaphic factor. Fieldwork has found that the main forests grow over partially acidic soils (in a mainly calcareous lithological environment in both massifs), allowing the species to tolerate the dryness of the summer period; this warm season is characterized by high sunshine and a notable absence of rainfall in July and August and also often in June and September, with less than 10 mm of rain in the driest month (Costa, et al., 2005). This partial development of *P. sylvestris* subsp. *nevadensis* on acidic soils means that the edaphic character of the series *Daphno-Pinetos nevadensis* must be redefined, and be considered as silicicolous and not only as limestone and dolomiticolous. In addition, due to its strong Mediterranean character it is predictable that the subsp. *nevadensis* is better adapted to the conditionings (warm, dryness and drought) of the Mediterranean climate of southern Iberian Peninsula than other northern Iberian varieties located in mountain areas that keep Eurosiberian ecological conditions. Martinez-Vilalta and Piñol (2002) and Höðar et al. (2003) point out that Spanish populations of *P. sylvestris* suffered high mortality rates associated with years of drought during 2005-2007; however, evidence has not been found to suggest that this climate event (with precipitation rates of between 25% and 75% below normal in the Betic Cordillera) has affected the subsp. *nevadensis*, with an abnormal death rate not having been detected.

However, the future is uncertain and the above hypothesis may prove unreliable if the forecasts that announce lower rainfall totals and more extreme periods of drought come true (Morata, 2014). This fact is more relevant considering that the Mediterranean Basin is one of the world’s most vulnerable regions to the global environmental threat, that it is already being affected by climate change (Giorgi, 2006), which will have numerous potential effects and will be a major driving factor in mountain ecosystem development in the next century (Jentschs, Beierkuhnlein, 2003; Löfler, et al., 2011). The alteration of conditions at a local level can be lethal for *P. sylvestris* subsp. *nevadensis*, which can be replaced by *P. nigra* (Benito, et al., 2006; Garzón, et al., 2008) due to the better adaptation of this species to the continental and relatively xeric conditions of the Betic Cordillera climate, something that is also expected to happen in other northern mountains of the Iberian Peninsula (Ruiz-Labourdette, et al., 2012). López and Camacho (2010) have estimated an 85% decline in the potential area of *P. sylvestris* in Spain, dropping from 84,447 km² under current climate conditions to only 12,691 km² under the predicted conditions in 2050 (based on the regionalised scenarios of climate change on Spain, Morata, 2014). Regarding the predicted warming in this area of the planet, Vandvik (2017) points out that temperature alone should not be used to predict the future evolution of tree seedling establishment above the alpine treeline, so extrapolations from climate envelope models could strongly over or under estimate the upper forest boundary responses to global change.

It is too early to accurately predict the future of *P. sylvestris* subsp. *nevadensis* because it is not possible to know the species current (and future) adaptation to adverse climatic conditions and global change, in particular in these mountain environments that are now subject to less anthropogenic alteration in recent decades, especially over the last 25 years as result of the new protected status of Sierra Nevada (Natural Park and National Park) and Sierra de Baza (Natural Park). The circumstances in which these forests evolve are just beginning to be understand. In this regard, future knowledge of these mountain areas should take into account the arguments of Löfler et al. (2011), that whilst many detailed studies have generated highly specific data related to the response mechanisms to global change in mountain areas, these results are often obtained under a narrow range of environmental conditions; in the study area, e.g., the climatic data collected above 2,000 m asl in Sierra Nevada are insufficient (only 20 years of observations at best) and, in some cases, unreliable, while in the upper part of
the Sierra de Baza there is no climatic information beyond data obtained from extrapolations, such as those collected in Olmedo-Cobo (2011). Thus, it is necessary for multi-factorial studies in order to understand the interplay between the processes of climatic change and other environmental factors, and their effects on tree seedling establishment across the upper boundaries of the current forest (Vandvik, 2017).

In this uncertain climatic scenario, the most negative factor for the current survival of the subsp. nevadensis in the geographical context of Betic mountain range is the long and complex process of natural regeneration of P. sylvestris in Mediterranean environments, which is not successful in all cases (González-Martínez, Bravo, 1999; Koski, 2000); especially at risk is recruitment at its southern most edge due to the predicted increase in summer drought (Escudero, et al., 1997; Martínez-Vilalta, Piñol, 2002; Castro, et al., 2004). Other serious obstacles to the conservation and development of seeds, seedlings, and juvenile individuals, which ensure forest renewal, are the frequent presence of wild ungulate herbivores (mainly Cervus elaphus L.) and livestock herds in the study area—the pressure from animals is significantly high during dry season—(Castro, 1999; Zamora, et al., 2001; Castro, et al., 2004, 2005). Other causes of mortality of young seedlings are insect attacks, small landslides on steep slopes, consumption of seeds by rodents, and hail storms or frost out of season (Castro, et al., 2004, 2005). Especially important is the attack of Thaumetopoea pityocampa D. & Schiff (pine processionary caterpillar); although it is worth noting that whilst this pest affects to a much greater extent the masses of conifers located between 800 and 1,500 m asl (Hódar, et al., 2003b), its effects over subsp. nevadensis are increasingly frequent as a result of climate change, and in spite of the isolation of these forests in high areas (Hódar, et al., 2003a). In fact, the authors of this investigation have verified how in winter of 2016 the pest has reached levels above 2,000 m asl affecting autochthonous masses of P. sylvestris (and P. nigra) of Sierra de Baza, so this factor may has contributed to the death of numerous trees already weakened by the drought that has affected this area in recent years (Figure 7 left) (Olmedo-Cobo, et al., 2016). In addition, this process of forest decay has also been aggravated by the presence—observed by the authors during spring and summer of 2016—of bark beetle (Tomicus genus, subfamily Scolytinae) causing a remarkable mortality of pine trees in low and medium zones of this massif (Figure 7 right), pest that threatens to also affect the forests of P. sylvestris subsp. nevadensis in the upper zone of the massif.

In addition, the reforestation carried out from the 1950s to the 1980s is another important obstacle for the regeneration and conservation of P. sylvestris subsp. nevadensis forests. This process is in part the result of the abandonment of rural Mediterranean moun-

![Figure 7. P. sylvestris subsp. nevadensis affected by T. pityocampa (left) and extensive masses of reforestation pines dead due to the action of Tomicus genus insects (right) in Sierra de Baza](source: Rodríguez-Sánchez)
of pollen gene flow in homogenizing populations on
this type of geographic scales (Pazouki, 2015), the
genetic conservation management plans should include
assisted regeneration and the maintenance of ex-situ
germlasm collections.

Finally, the communities as a whole that comprise
the phytocoenosis preserve numerous endemic spe-
cies, some of which (including *P. sylvestris* subsp. *nevadensis*) are among the most endangered Mediter-
nanean flora, especially those with a small chorological
area that only appear in the Betic biogeographical
province. Therefore, these forests have a “protective”
role that favours the growth and conservation of part
of these taxa. Due to all these circumstances, Sier-
ra Nevada and Sierra de Baza can be included in the
Mediterranean core refuges of plants in which the
preservation of the most endangered species should
be a priority to ensure their future persistence and ge-
netic diversity (Médail, Diadema, 2009). These pro-
tectionary needs have been also remarked on by sev-
eral other pieces of research exploring other relict
peri-Mediterranean mountain and endangered spe-
cies in Betic Cordillera, such as *Abies pinsapo* Boiss.
by Olmedo-Cobo et al. (2017) and Gómez-Zotano et
al. (2014), and *Taxus baccata* L. by Olmedo-Cobo and

**Conclusion**
The exceptional and relict character of *P. sylvestris*
subsp. *nevadensis* endemic forests is related both with
the physical characteristics of the microenvironments
in which the subspecies has taken refuge, as with the
small size of its populations, the lack of regeneration
and the absence of young specimens due to overgraz-
ing. The cool, relatively moist climate in the high-
er areas of the Betic Cordillera, in particular in Sier-
rar Nevada and Sierra de Baza massifs, the favourable
N orientation and exposure to more humid air mass-
es from the W, and the relatively impermeable soils
that the forests occupy are the main natural factors
that have allowed the survival of *P. sylvestris* in the
southern Iberian Peninsula under hostile Mediterrane-
nan environment conditions. Additionally, the natu-
ral process of regression and cantonment of the subsp.
*nevadensis* forests in these disjunct and refuge areas,
inside favourable ecological niches of upper treeline
boundary, has been accelerated and strengthened in
the last millennia by the intense anthropo-zoogenic
disturbance to which these mountains have been sub-
jected, within the broader context that has affected
the Mediterranean basin. Thus, the human factor has
decisively contributed to the uniqueness, rarity and
endangered status of *P. sylvestris* subsp. *nevadensis* at
present.

Given these circumstances, these forests can be
considered as post-glacial relics well adapted to the
singular ecological Mediterranean conditions exist-
ing in the upper boundary treeline of Sierra Nevada
and Sierra de Baza. This fact is especially important
given the vulnerability of the Mediterranean Basin
to predicted environmental changes; if the trend to
warmer, drier climate conditions continues, this type
of phytocoenosis will be the first to be in serious dan-
ger. Therefore, the protection of these high mountain
Mediterranean forests threatened with extinction is
particularly important.

In order to design meaningful plans for protecting
these relict ecosystems, it is necessary to start solid
programmes for monitoring the evolution of *P. sylves-
tris* subsp. *nevadensis* and studying the processes that
make this species so unique and valuable. This also
requires a broader knowledge of the ecological role
of the biodiversity in the territory, because a less plu-
ralistic area-based strategy (e.g. flora micro-reserves)
for isolated species that are not strongly functionally
connected to other taxa would be more effective than
a National or Natural Park. Nonetheless, the declara-
tion in past decades of these massifs as protected are-
as has provided a necessary, ideal, and unique support
for the study, the management, and the protection of
the *P. sylvestris* subsp. *nevadensis* relict forests in
southern Spain.

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paper.

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Pinus sylvestris L. subsp. nevadensis (Christ) Heywood in southern Spain: An endangered endemic Mediterranean forest


changes in a Quaternary refugium: evolutionary